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Dams, Malaria, and Agriculture
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Abstract

We study the effect of large dams on malaria incidence in India between 1975 and 1995. We

combine an instrumental variables approach with a panel model with unobserved common factors

allowing to fully capture the endogeneity of dam location and unobserved time-varying

heterogeneity. Dams result in increased malaria incidence in districts where dams are located and

in downstream areas. We find that the construction of a large dam increases a district's annual

malaria incidence by about 0.9 to 1.4 per cent, and by about 1 to 1.5 per cent in downstream

districts. We also find that this malaria-increasing effect of dams persists over time. Our results

imply that the construction of dams in malaria-sensitive areas should be coupled with direct

interventions, such as the wide deployment of insecticide-treated nets or the roll-out of future

vaccines. Finally, we examine the contribution of agricultural development to this malaria-

increasing effect of dams. We find that dam construction benefits agriculture in the vicinity of

dams, as well as in downstream areas. These positive effects are driven by increased irrigation

and cultivation in the vicinity of dams, while they are driven by changes in cropping patterns in

downstream areas, where the cultivation of HYV crops increases.

Keywords: irrigation, large dams, malaria, agriculture, India, SDG 3.3.

JEL Codes: Q19; O20.

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"When he builts [sic] these [dams], Man plays the role of Sorcerer's Apprentice"

- Rothé, 1968.

Introduction

Large irrigation dams play a vital role in the implementation of several key Sustainable Development Goals (SDG), because they potentially provide a range of benefits, from preventing river floods, improving access to electricity and reliable clean water, strengthening the resilience of agricultural systems in the face of climate change, reducing poverty and food insecurity, to promoting economic and social development (e.g. Duflo and Pande, 2007; Kibret, 2018; Dillon and Fishman, 2019).

However, dams have been criticized for several reasons. For instance, the geographical distribution of (most of) these benefits may be unequal in that the gains associated with dams may largely benefit the command area, typically downstream, while the costs might be concentrated in areas where dams are located, or the catchment area (e.g. Duflo and Pande, 2007). Dams might also come with a range of negative environmental impacts, e.g. increased silting, salinity, or pollution risks (e.g. Mettetal, 2019).

In addition, dams may result in pervasive health effects, especially linked to risks of water-related infectious diseases (e.g. Kibret, 2018; Mettetal, 2019). In particular, an increasing body of anecdotal reports and case studies suggest large dams are associated with increased malaria cases and deaths (e.g. Yewhalaw et al., 2013; Dejenie et al., 2012).

Malaria is a life-threatening disease caused by parasites that are transmitted to people through the bites of infected female Anopheles mosquitoes. Dams may indeed be ideal habitats for

the breeding and egg-laying of these mosquitoes as their shorelines provide many shallow water puddles and pools where predators to mosquitoes may be absent (e.g. fish or tadpoles) (e.g. Kibret, 2018).

While more than 200 million people are affected by malaria worldwide¹, the World Health Organization aims at reducing global malaria cases and mortality rates by at least 90 per cent by 2030. The push to end malaria is also recognized in the Sustainable Development Goals Target 3.3 "By 2030, end the epidemics of AIDS, tuberculosis, malaria and neglected tropical diseases and combat hepatitis, water-borne diseases and other communicable diseases". Given this policy background, this paper examines the impact of large dams on malaria incidence.

In the face of accumulating anecdotal evidence or descriptive analyses suggesting dams may be associated with malaria (e.g. Lautze et al., 2007; Kibret et al., 2019), several systematic reviews have, however, concluded otherwise (e.g. Dillon and Fishman, 2019; Keiser et al., 2005). For example, Keiser et al. (2005) conclude that "no clear relationship emerges from this literature review between the risk factor "development and operation of water projects" [sic] and malaria risk". Dillon and Fishman (2019) suggest that "the pathway of higher malaria incidence related to increased hydrological investments is not well substantiated". These reviews also highlight the dearth of evidence in the empirical literature and call for additional research².

An important limitation of the literature is that existing studies remain vulnerable to several biases, among which is the endogeneity bias due to the non-random nature of dam location, and thus only provide non-causal evidence (e.g. Duflo and Pande, 2007; Dillon and Fishman, 2019)³.

¹ While the largest share of the malaria burden falls on Africa, the majority of people living in close proximity of the reservoirs of large dams in malaria-endemic areas are in India.

² The case for reliable research on large dams is even more compelling because of their controversial nature (e.g. Duflo and Pande, 2007; Strobl and Strobl, 2011; Dillon and Fishman, 2019).

³ The non-random location of dams may confound estimates. For example, local governments may target areas whose agriculture is either already productive.

To establish potential causality between dams and malaria is indeed challenging from an econometric standpoint, and virtually all existing studies have failed to do so. A notable exception is Duflo and Pande (2007) – DP thereafter – who provide a viable causal identification strategy by relying on various geographical features of the district. To be more specific, they exploit variation in dam construction induced by differences in river gradient across districts within Indian states to construct instrumental variable estimates and find no effect of large dams on malaria incidence in India.

Still, one issue with DP is that they rely on state-year fixed effects to partially account for time-varying omitted variables. Doing so imposes that the impact of unobservable common shocks is homogenous on malaria across all districts within a state. This is unlikely to be valid for several reasons because state-level common shocks, for example, climate shocks or input/output price shocks, would likely have differentiated effects on malaria incidence across districts within a state, dependent on the district's topology, type of agricultural practices and systems, health systems and infrastructures, and existing mitigation and adaptation capacities (e.g. Skoufias, 2003; Kuriakose et al., 2013; Asfaw et al., 2017).

More generally, there is no reason to assume that the impact of such common shocks would not be heterogeneous across districts. Assuming otherwise may result in biased estimates. For example, Mary (2022) shows that restrictive identification assumptions with respect to modelling time dependence, including the use of state-year fixed effects, may create an omitted variable bias and lead to largely misguided policy implications in the case of the mitigating impact of dams on ethnic conflict in the face of rainfall shocks in India.

Recently, models with unobserved common factors or interactive fixed effects have been increasingly proposed to solve similar issues in the general economics literature. Such approach

captures time-varying heterogeneity via a set of factor loadings and requires the imposition of a factor structure. While common factors or interactive effects models are somewhat more complicated than fixed effects models, they have increasingly become commonplace in the empirical economics literature (e.g. Gobillon and Magnac, 2016).

The literature has provided several modeling approaches to estimate panel models with unobserved common factors or interactive effects: common correlated effects (CCE) (Pesaran, 2006), interactive principal components (IPC) (Bai, 2009; Bai and Li, 2014), and general two-stage IV (G2SIV) approaches (Norkute et al., 2021; Cui et al., 2022). All these approaches have benefits and drawbacks. However, the latter may be a preferred option because it combines the attractive features of both the CCE and IPC approaches (Kripfganz et al., 2021).

In this paper, we build on Duflo and Pande (2007) and combine their instrumental variables (IV) approach with the G2SIV approach for estimating panel models with unobserved common factors from Norkute et al. (2021) and Cui et al. (2022) to control for, respectively, the endogeneity of dam placement, and time-varying unobserved heterogeneity. Given our modeling approach, we assess the effect of technologically feasible dams on malaria.

Our paper provides evidence that large irrigation dams increase malaria incidence. We find that dams increase malaria incidence in districts where dams are located, and in downstream districts. Our estimates imply a one-dam increase in the number of dams in the district increases annual malaria incidence by about 0.9 to 1.4 per cent on average. The effect in downstream districts is about 1 to 1.5 per cent. This confirms the multiple anecdotal and descriptive reports that have suggested a potential link between dams and malaria (e.g. Kibret et al., 2019; Lautze et al., 2007; Ghebreyesus et al., 1999).

To understand the mechanisms linking dams and malaria, we provide a second set of agriculture-focused results. These results suggest that dam construction benefits agriculture in the vicinity of dams, as well as in downstream areas. These positive effects are driven by increased irrigation and cultivation in the vicinity of dams, while they are driven by changes in cropping patterns in downstream areas. In particular, the cultivation of HYV crops, which require more water, increases in downstream areas. These agricultural results highlight the potential mechanisms between dam construction, irrigation area, cropping patterns, and malaria incidence.

Overall, we contribute to the literature on irrigation dams in three important ways. First, from an empirical standpoint, we show that large dams affect malaria incidence in India. Second, we show that dam construction benefits agriculture in the vicinity of dams, as well as in downstream areas. These findings are novel in the empirical literature and will contribute to a reassessment of the distributional cost-benefit analyses linked to the construction of large dam infrastructures from both health and economic perspectives. Third, from a methodological perspective, we highlight the role of time-varying heterogeneity in empirical studies examining the impact of irrigation infrastructures, the limitation of state-year fixed effects to account for unobserved time-varying heterogeneity in this context, and finally demonstrate the usefulness of common factors approaches.

The rest of paper is structured as follows. Section 2 provides some background on the links between dams and malaria. Section 3 describes the identification strategy. Section 4 presents the data. Section 5 analyses the empirical results while Section 6 concludes.

Background

Large dam construction has been increasingly met with strong opposition from ecologists and environmentalist groups. Arguments from these opposition groups usually reference the ability of large dams to fragment rivers and disrupt their flows, which threatens ecosystem services that people and wildlife depend on for their survival (McCartney and Sally 2007; FitzHugh and Vogel 2011; Wildi 2010). Furthermore, research into the health impacts of large dams has also heightened through recent case studies and anecdotal evidence (e.g. Dillon and Fishman, 2019). This line of research particularly focuses on the impact that large dams have on food and water scarcity, communicable diseases, social disruption, and vector-borne diseases such as schistosomiasis and malaria (e.g. Lerer and Scudder, 1999). This paper focuses on the causal link between large dams and malaria incidence.

Once constructed, large dams can create small pools of standing water, which is an ideal breeding ground for adult female Anopheles mosquitoes, the primary vector of malaria transmission (International Water Management Institute 2018; Sadoff 2018). These small pools of water, often caused by flooding within the vicinity and upstream of the dam, create a year-round habitat for Anopheles mosquitoes, extending the period of malaria transmission in some regions to include both the rainy and dry seasons (Mbapok et al., 2018). Furthermore, the slope of the dam can also affect the Anopheles population in the area. In fact, dams with gentle and shallow slopes allow pools of water to easily form, which further increases the probability of Anopheles habitat formation.

Conversely, large dams can disrupt the formation of Anopheles breeding habitats that are located downstream of the dam by reducing river flows and decreasing the amount of water discharge in the area (Lee et al., 2019; Larinier, 2001). This reduction in the water volume downstream of the dam limits the possibility of Anopheles habitat formation, which can greatly

reduce the transmission of malaria in the area. This effect is especially prominent during the dry season.

Large dams can also exacerbate the impact that agricultural development has on malaria transmission in a variety of ways (e.g. Janko et al., 2018). First, dams can greatly expand the role of irrigation as an agriculture practice in nearby farmlands. If these irrigation systems lack efficient drainage, water can accumulate into small standing pools, providing additional breeding ground for Anopheles mosquitoes (GBCHealth, 2012). Other mechanisms in which dams can facilitate an increase in malaria incidence through agriculture expansion include deforestation, increased crop cover, and agriculture encroachment in highland regions (Asenso-Okyere et al., 2009; MacDonald and Mordecai 2010; Guerra et al., 2006). Finally, migrants from malaria endemic areas seeking work or food and water security can travel to these newly developed farmlands, which could also greatly facilitate the transmission of malaria (Falavigna-Guilherme et al., 2005)

Against this conceptual background, the empirical literature remains limited, provides mixed findings, and mostly consists of case studies and anecdotal evidence. Lautze et al. (2007) analyze the impact that the Koka dam had on malaria incidence in the Rift Valley of Ethiopia from 1994 to 2002. They find that malaria case rates among people living within 3 km of the Koka reservoir were 1.5 times greater than those living between 3 and 6 km from the reservoir and 2.3 times greater than those living 6-9 km from the reservoir. Additionally, Ghebreyesus et al. (1999) find that the incidence of malaria is significantly higher in children who live in villages close to dams (within 3 km) than in children who live further away (between 8-10 km). Recently, Kibret et al. (2019) find that distance to reservoir shoreline and malaria incidence were negatively correlated in their analysis of three dams in Ethiopia⁴ from 2010-2014.

⁴ Koka dam, Kesem dam, Koga dam.

In contrast, Yewhalaw et al. (2013) study the impacts that an Ethiopian dam, the Gilgel-Gibe dam, had on malaria incidence from 2008 to 2010 but find no statistically significant relationship between malaria incidence and proximity to dam. However, the authors do find that the malaria vector (Anopheles arbiensis) was more abundant in villages closer to the reservoir. Likewise, in their study of six dams in Ethiopia's Tigray Region, Dejenie et al. (2012) find that malaria prevalence did not increase in villages within 4 km of a dam compared to villages further away⁵.

Ndiath et al. (2012) study the impact of agriculture development on malaria transmission in the Senegal River basin after the construction of two dams along the Senegal River. They find that malaria transmission was low and seasonal during their two-year study period. Interestingly, the authors also find that Anopheles densities were significantly correlated with the presence of ditch water (for activities like animal watering and brick manufacturing) but not correlated with proximity to the river. Furthermore, Janko et al. (2018) find that rural children under the age of 5 in the Democratic Republic of Congo were particularly vulnerable to malaria transmission resulting from similar agriculture development. Falavigna-Guilherme et al. (2005) find that the construction of the Itaipu Dam on the Paraná River resulted in the proliferation of the Anopheles mosquito in the area and the subsequent malaria outbreak.

In their extensive analysis on the impact dams have on malaria incidence in sub-Saharan Africa, Kibret et al. (2019) suggest that between 0.7 and 1.6 million malaria cases per year were attributable to large dams from 2000-2015. Further, in their univariate correlation analysis, the authors find that the slope of the reservoir explained 46.8 per cent of the variation in malaria incidence around large dams, more than any other environmental factors, including rainfall,

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⁵ However, they also report that the population of adult Anopheles mosquitoes increased in villages closer to a dam than those further away.

temperature, and humidity. They also find that the slope of a reservoir significantly impacted the malaria incidence in communities less than 5 km from a dam.

Unfortunately, these descriptive studies cannot provide causal evidence of a link between dams and malaria incidence because they do not account for the endogeneity bias that occurs from the non-random allocation of dams. A notable exception to the literature is the analysis from Duflo and Pande (2007) who study the impact dams have on agricultural production and rural welfare in India. To control for the endogeneity issue, they use differences in river gradients across districts in Indian states to construct instrumental variables estimates. The authors find no evidence that dams affect malaria incidence.

However, Duflo and Pande (2007) do not fully account for time-varying heterogeneity as they only include state-year fixed effects in their analysis. As explained in the introduction, this modeling approach assumes that the impact of unobservable common shocks, such as climate shocks or agricultural input/output price shocks, is homogenous across all districts within a state. This is unlikely considering that these districts have varying agriculture practices, topology, health systems and practices, and adaptation policies that could impact their response to such shocks.

Overall, due to the regional focus, the somewhat mixed results, and existing methodological limitations in the literature, there is little convincing evidence that dams cause malaria incidence. In fact, extensive systematic reviews on the literature also confirm the lack of evidence (e.g. Keiser et al., 2005). For example, Keiser et al. (2005) did not find any conclusive evidence that dams impact malaria incidence in their systematic literature review. Another literature review by Dillon and Fishman (2019) suggests that there is little support for the link between higher malaria incidence and increased hydrological investments (Dillon and Fishman,

2019). As a result, further research is required to better inform policymakers and researchers on the impact of dams on malaria incidence.

Empirical strategy

Empirical model

To examine the impact of dams on malaria incidence, we can start with the following model that has been used in previous studies:

$$y_{ist} = \beta D_{ist} + \alpha D_{ist}^{U} + \gamma Z_{ist} + \delta Z_{ist}^{U} + \chi C_{ist} + \theta_i + \rho_{st} + \varepsilon_{ist}$$
 (1)

Where y_{ist} is the outcome of interest, i.e. malaria incidence, in district i in state s at time t; D_{ist} is the number of dams in the district; D_{ist}^U is the number of dams located upstream of the district; Z_{ist} and Z_{ist}^U is a vector of controls for the district and for upstream districts. In an extended model, C_{ist} is vector of climatic variables, namely rainfall and temperature shocks. θ_i are district fixed effects, controlling for time-invariant characteristics. θ_{st} are state-year (interacted) fixed effects partially controlling for time-varying heterogeneity with homogenous common shocks across districts in a state. β captures the impact of a dam on malaria in the district where the dam is located (i.e. the own district effect), while α describes the downstream impact of a dam.

There are two main problems with this model. First, the literature has recognized that dam construction may not be random. Indeed, local governments may target areas that are already more productive (but *perhaps* associated with higher malaria burden⁶) or politically better-connected

⁶ It could also be lower as wealthier districts may be better prepared to fight the disease with larger human, financial, and medical, infrastructures and resources.

(Mettetal, 2019). By contrast, the construction of dams may be seen as an engine of growth in poorer districts.

To control for this endogeneity issue, several studies have used the IV strategy designed by Duflo and Pande (2007)⁷. They exploit variation in dam construction induced by differences in river gradient across districts within states to construct IV estimates. Following their approach, we estimate:

$$D_{ist} = \alpha + \sum_{k=2}^{4} \varphi_k (RG_{ki} * \overline{D}_{st}) + \omega (M_i * \overline{D}_{st}) + \sum_{k=2}^{4} \pi_k (RG_{ki} * l_t) + \theta_i + \rho_{st} + \varepsilon_{ist} \quad (2)$$

Where D_{ist} is the number of dams in the district; RG_{ki} are the river gradient variables; \overline{D}_{st} is predicted dam incidence in the state; M_i is a vector of district-specific time invariant variables, i.e. district elevation, overall gradient measures, river length, and district area. Equation (2) also includes district fixed effects and state-year fixed effects. The interaction of river gradient variables with country-level common shocks $RG_{ki} * l_t$ accounts for national-time varying effects of river gradient on the outcomes of interest. The river gradients capture the proportion of the river with a gradient between 1.5 and 3%, 3 and 6%, and above 6%.

The parameters from Equation (2) can be used to predict the number of dams in the district, $\widehat{D_{lst}}$, that is used as an instrument for D_{ist} . DP also rely on Equation (2) to predict the number of dams upstream, arguing that their approach is more efficient because it uses all available variation to predict the relationship between river geographic features and the number of dams (rather than only those upstream). If a district does not have an upstream district, $\widehat{D_{lst}^U}$ is replaced with 0.

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⁷ This strategy (or a variant of) has been used implicitly or explicitly in Strobl and Strobl (2011), Sarsons (2015), Blanc and Strobl (2014) or Mettetal (2019).

The next step of the empirical strategy comes down to estimating a marginally modified version of Equation (1):

$$y_{ist} = \beta \widehat{D_{ist}} + \alpha \widehat{D_{ist}}^{U} + \gamma Z_{ist} + \delta Z_{ist}^{U} + \chi C_{ist} + \theta_i + \rho_{st} + \varepsilon_{ist}$$
 (3)

Where Z_{ist} and Z_{ist}^U include, respectively, all controls from Equation (2) but the interacted river gradient variables with predicted dam incidence in the state. Estimating Equation (3) captures the local average treatment effect (LATE) of technologically feasible dams (or in other words, the LATE of dams constructed for geographical reasons).

The second problem, which has been much less discussed in the literature, relates to how unaccounted time-varying heterogeneity (TVH) may affect estimates (e.g. Blanc and Strobl, 2014). To control for TVH, the literature related to dams has used a variety of fixes. Duflo and Pande (2007) use state-year fixed effects. Sarsons (2015) uses state-level time trends. Mettetal (2019) uses a combination of year fixed effects and district-level trends.

In the face of such diverse modeling choices, it is critical to understand their implications. Mary (2022) shows that restrictive assumptions with respect to modeling time dependence to partially capture time-varying heterogeneity may result in largely biased findings in the case of the links between rainfall shocks, dams, and ethnic conflict in India. Therefore, we argue that state-year fixed effects are not sufficient to fully address concerns about time-varying heterogeneity.

Ideally, one would like to replace state-year fixed effects with district-year fixed effects. This is of course in practice impossible due to the incidental parameter problem leading to biased and inconsistent estimates. To circumvent this challenge, common factor approaches⁸ have been

⁸ To validate the use of factor models, we use the cross section dependence test from Pesaran (2015). The hypothesis of null cross section dependence is rejected at the 1 per cent level in our application. This suggests that the presence of common factors influencing malaria across districts and explicitly validates the factor model approach.

increasingly used to account for rich sources of unobserved heterogeneity. Following the introduction of common factor approaches, we can modify Equation (3):

$$y_{ist} = \beta \widehat{D_{ist}} + \alpha \widehat{D_{ist}}^{U} + \gamma Z_{ist} + \delta Z_{ist}^{U} + \chi C_{ist} + \theta_i + \rho_{st} + \mu_i F_{st} + \varepsilon_{ist}$$
(4)

Where F_{st} is a set of common factors and μ_i a vector of factor loadings. μ_i represents the heterogeneous impact of unobservable common shocks within a state on each district. Let's finally note that estimating Equation (4) is exactly equivalent to estimating Equation (3) if common shocks have homogenous effects on the outcome of interest across districts within a state, $\mu_i = \mu$.

Furthermore, if we ignore the endogeneity of dam location, we estimate a modified version of Equation (4) where the numbers of dams in the district and upstream are not instrumented:

$$y_{ist} = \beta D_{ist} + \alpha D_{ist}^{U} + \gamma Z_{ist} + \delta Z_{ist}^{U} + \chi C_{ist} + \theta_i + \rho_{st} + \mu_i F_{st} + \varepsilon_{ist}$$
 (5)

In other words, this latter model accounts for time-varying heterogeneity but ignores the non-random allocation of dams. The estimations of Equations (4) and (5), in essence, provide respectively IV- and OLS-like estimates in a common factors framework.

Last, we transform the models in Equations (4) and (5) by introducing a lagged dependent variable to account for the existence of persistence in malaria. We obtain, in the case of Equation (4):

$$y_{ist} = \sigma y_{ist-1} + \beta \widehat{D_{ist}} + \alpha \widehat{D_{ist}} + \gamma Z_{ist} + \delta Z_{ist}^{U} + \chi C_{ist} + \theta_i + \rho_{st} + \mu_i F_{st} + \varepsilon_{ist}$$
 (6)

Where σ captures persistence in malaria cases. In the case of Equation (5), we get:

$$y_{ist} = \sigma y_{ist-1} + \beta D_{ist} + \alpha D_{ist}^{U} + \gamma Z_{ist} + \delta Z_{ist}^{U} + \chi C_{ist} + \theta_i + \rho_{st} + \mu_i F_{st} + \varepsilon_{ist} (7)$$

Estimation

The literature has provided several modeling approaches to estimate panel models with unobserved common factors or interactive effects: common correlated effects (CCE) (Pesaran, 2006), interactive principal components (IPC) (Bai, 2009; Bai and Li, 2014), and general two-stage IV (G2SIV) approaches (Norkute et al., 2021; Cui et al., 2022).

All these approaches have benefits and drawbacks. However, the latter may be a preferred option because it combines the attractive features of both the CCE and IPC approaches (Kripfganz et al., 2021). First, unlike CCE or IPC, the G2SIV approach does not suffer from incidental parameters bias and thus does not require bias correction. Second, the IV approach is robust and computationally inexpensive. Third, the approach allows for endogenous regressors (if external instruments are available). Fourth, it can be applied with unbalanced panels.

In a first stage, the IV approach⁹ projects out the common factors from exogenous covariates using principal components analysis and constructs instruments from defactored covariates. In a second stage, the entire model is defactored based on factors extracted from the first-stage residuals, and IV regression is implemented again using the same instruments. This approach removes the common factors form the error term and the regressors separately in two stages.

There are a few practical choices to be made for the implementation of the IV approach. We mainly rely on the Ahn and Horenstein (2013) eigenvalue ratio test to compute the number of factors jointly. We set the maximum number of factors to 4 for each estimation stage and each set of instruments. Finally, note we double down on our common factors approach and also introduce

 $^{^9}$ We use the Stata command xtivdfreg designed by Kripfganz et al. (2021) to implement the approach, and the command xtnumfac from Ditzen and Reese (2022) .

common factors in Equation (2) to make the entire estimation strategy consistent. We also provide supporting evidence for why this is needed given the data.

Data

Data is mainly taken from the dataset compiled by DP that is publicly available at https://dataverse.harvard.edu/dataset.xhtml?persistentId=hdl:1902.1/IOJHHXOOLZ. The dataset we use includes information on district geography, dam placement, river gradient, malaria incidence, between 1975 and 1995 for a sample of 358 Indian districts. While we refer the reader to DP for more details, we summarize key information about each variable below. Descriptive statistics can be found in Table A1 in Appendix.

Data on dams come from the World Registry of Large Dams. A large dam is defined as a dam having a height of 15 meters from the foundation, or, if the height is between 5 and 15 meters, having a reservoir capacity of more than 3 million cubic meters. The registry lists the nearest city to the dam and allows for allocating dam to district.

The malaria incidence data used was collected by India's National Malaria Eradication Program (NMEP). Malaria incidence is measured as the log of annual parasite incidence (API), where API is defined as the number of smears positive for *P. faliciparum* out of the population under NMEP's surveillance. Malaria data is only available from 1975 and 1995.

Rainfall data comes from the University of Delaware's Center for Climatic Research, which provides rainfall estimates at 0.5° longitude and latitude nodes across the world. A rainfall shock is measured as the fractional deviation of rainfall from its average level (calculated over 1916-1995), summed over all months. We also add temperature data to this dataset. Data are taken from https://www.indiawaterportal.com. This data is derived from the publicly available Climate

Research Unit TS2.1 dataset from the Tyndall Centre for Climate Change Research, School of Environmental Sciences, University of East Anglia in Norwich, UK. Temperature shocks are defined using the same methodology described above for rainfall shocks. Data on district area, river kilometers, elevation, overall gradient, river gradients come from CIESIN, Earth Institute Columbia University.

As in previous studies relying on India data (e.g. Sarsons, 2015; Duflo and Pande, 2007), the unit of analysis is the district as the most disaggregated data is only available at this level. A district is the administrative unit immediately under a state (we use the Census 1981 as reference). Duflo and Pande (2007) argue that the district in which a dam is located includes most or all of its catchment area, and a portion of the command area, while most of the command area is in the downstream district. They suggest that the effect of a dam will be unambiguously positive downstream (and likely underestimated). By contrast, it is not clear what to expect overall for the district in which a dam is located (as the positive 'command' effects could dominate the negative 'catchment' effects, or vice versa).

Empirical results

The importance of geography in large dam construction

Table 1 displays the G2SIV estimation results of Equation (2) in column 1, where the number of dams in the district is the dependent variable and regressed against river gradients and a set of geographical variables that are all interacted with the predicted dam incidence at the state level. District fixed effects, state-year fixed effects, and common factors are included. The estimation includes 358 districts between 1975 and 1995.

Table 1. The importance of geography in large dam construction

	Interacted with predicted number of dams in the state
Number of dams	(1)
Fraction river gradient 1.5-3%	11.619***
	(4.117)
Fraction river gradient 3-6%	-12.671*
	(7.622)
Fraction river gradient above 6%	11.785***
	(2.436)
River length	-0.000
	(0.001)
District area	0.000***
	(0.000)
Elevation 250-500m	-0.414
	(1.025)
Elevation 500-1000m	2.881*
	(1.484)
Elevation above 1000m	-4.771
	(5.661)
District gradient 1.5-3%	10.500*
	(6.073)
District gradient 3-6%	3.528
	(9.504)
District gradient above 6%	1.126
	(5.792)
F-test for river gradient, p-val. (stat.)	0.000 (97.38)
N	6937
Number of districts	358
Common factors	YES
District FE	YES
Rho	0.763
J-test	0.53
Number of instruments	44
Number of common factors	1
Number of lags	3 yel <0.01 ** p.yel <0.05 * p.yel <0.1 # p.yel <0.15 All

Notes: Robust standard errors in parentheses. *** p-val. <0.01, ** p-val. <0.05, * p-val. <0.1, # p-val. < 0.15. All regressors are interacted with predicted dam incidence at the state level. Rho is the fraction of the variance of the error term that is explained by the factor component. Note these estimations are based on a full sample ranging from 1975 to 1995. All G2SIV estimates. The sample includes 358 districts.

All covariates are defactored based on a common set of factors estimated jointly. We use 3 lags of all covariates as defactored instruments, up to a maximum of 1 factor following the Ahn and Horenstein (2013) criteria. We control for district-specific and state-year fixed effects by eliminating them prior to the estimation. The p-value of the J test statistic (0.53) indicates that the over-identifying restrictions (instruments) are valid. The estimated number of factors in the first and second steps both equal 1.

The rho is about 0.76 in column 1, suggesting that most of the variation in the composite error term is due to the single unobserved factor, conditional on district fixed effects, and state-year fixed effects. Put simply, estimators that do not control for common shocks are likely to be severely biased. In essence, our results in column 1 mimic those of Duflo and Pande (2007) and support the importance of engineering considerations for dam construction. We find that a gentle river gradient (1.5-3%) and a very steep river gradient (3-6%) increase the number of dams. Interestingly, a steep river gradient (3-6%) has a negative effect on dam construction. The p-value for the F-test for the joint significance of all river gradient variables is 0.00 (F-stat: 97.38). In addition, district area seems to be a positive factor in building dams in upstream districts.

The impact of large dams on malaria incidence

Table 2 displays G2SIV estimates of the impact of large dams on malaria incidence both in the district where they are located ("Dams") and downstream ("Upstream dams") for the baseline model. Columns 1, 2, and 3 respectively display G2SIV estimates where the number of common factors changes from 3 to 1. We rely on several selection criteria to choose the number of common factors and they suggest this number to be either 1 (based on Ahn and Horenstein, 2013) or 3 (based on Onatski, 2010).

In our baseline specification, the two dam variables are defactored based on a common set of factors estimated separately, based on its own estimated factors, from the one for all the other variables. We use 6 lags of covariates as defactored instruments, and up to a maximum of 3 factors. The number of lags is chosen to minimize the variance of the model's standard error. We control for district-specific and state-year fixed effects by eliminating them prior to the estimation. In our baseline model, the estimated number of factors in the first step equals 1 throughout columns. In robustness analyses, we test whether our main results hold to alternative specifications. We discuss these results later.

Finally, as explained earlier, all columns use the predicted number of dams (in own district and downstream) as instruments for the actual number of dams. The J test statistic in column 1 does not reject the null hypothesis as the p-value equals 0.582. This implies that the additional instruments, i.e. predicted numbers of dams in own district and upstream, may be valid instruments.

The rho across columns in Table 2 is between 0.2-0.4, suggesting that a substantial part of the variation in the composite error term is due to the single unobserved factor, conditional on district fixed effects and state-year fixed effects. Put simply, estimators that do not control for common shocks may be likely biased.

The main message from Table 2 is that an increase in the number of dams in a district increases malaria incidence both in the district and in downstream areas. Our baseline estimate in column 1 implies that one-dam increase in a district would increase malaria incidence by about 0.9 per cent. Overall, according to columns (1)-(3), a one-dam increase in a district would increase malaria incidence by between 0.9 to 1.4 per cent. The construction of a dam in the district likely

creates additional breeding grounds for Anopheles mosquitoes that facilitate the development of malaria transmission.

Table 2. Baseline results

Malaria, log	(1)	(2)	(3)
Malaria, log, t-1	0.654***	0.523***	0.524***
	(0.006)	(0.007)	(0.009)
Dams	0.913***	1.096***	1.461***
	(0.301)	(0.349)	(0.401)
Upstream dams	1.085***	1.462***	1.041**
	(0.309)	(0.384)	(0.430)
N	2,845	2,845	2,845
Number of districts	355	355	355
Geographical controls	YES	YES	YES
Common factors	YES	YES	YES
District FE	YES	YES	YES
Rho	0.396	0.355	0.213
J test	0.582	0.429	0.356
Number of instruments	140	140	140
Number of common factors	3	2	1
Number of lags	6	6	6

Notes: Robust standard errors in parentheses. *** p-val. <0.01, ** p-val. <0.05, * p-val. <0.1, # p-val. < 0.15. Dams and upstream dams coefficients are multiplied by 100. Rho is the fraction of the variance of the error term that is explained by the factor component. Note these estimations are based on a sample ranging from 1975 to 1995. All G2SIV estimates. The sample includes 355 districts, instead of 358, due to missing data. All geographical variables (interacted with predicted dam incidence at the state level) from Table 1 are included except for the river gradient variables.

Furthermore, we find some evidence that dams also increase malaria incidence in downstream districts in all columns. According to columns (1)-(3), a one-dam increase in a district would increase malaria incidence by between 1 to 1.5 per cent. We assume this could be due to agricultural development in downstream areas. We will examine potential agricultural mechanisms later in the paper.

Table 3. Changing the number of lags used in the G2SIV estimation

Malaria, log	(1)	(2)	(3)	(4)	(5)	(6)
Malaria, log, t-1	0.713***	0.702***	0.606***	0.605***	0.536***	0.539***
	(0.021)	(0.022)	(0.013)	(0.013)	(0.2165)	(0.001)
Dams	0.883***	1.374***	0.892**	0.945**	1.284***	1.541***
	(0.293)	(0.456)	(0.436)	(0.444)	(0.289)	(0.215)
Upstream dams	0.195	-0.658#	0.443	0.332	1.135***	1.620***
	(0.416)	(0.420)	(0.406)	(0.375)	(0.263)	(0.237)
N	4,177	4,177	3508	3,508	2186	2186
Number of districts	357	357	356	356	352	352
Geographical controls	YES	YES	YES	YES	YES	YES
Common factors	YES	YES	YES	YES	YES	YES
District FE	YES	YES	YES	YES	YES	YES
Rho	0.142	0.350	0.166	0.396	0.221	0.467
J test	0.218	0.386	0.046	0.040	0.435	0.530
Number of instruments	100	100	120	120	160	160
Number of common factors	1	3	1	3	1	3
Number of lags	4	4	5	5	7	7

Notes: Robust standard errors in parentheses. *** p-val. <0.01, ** p-val. <0.05, * p-val. <0.1, # p-val. < 0.15. Dams and upstream dams coefficients are multiplied by 100. Rho is the fraction of the variance of the error term that is explained by the factor component. Note these estimations are based on a sample ranging from 1975 to 1995. All G2SIV estimates. The sample includes 355 districts, instead of 358, due to missing data. All geographical variables (interacted with predicted dam incidence at the state level) from Table 1 are included except for the river gradient variables.

These findings are the first evidence of the link between dam and malaria incidence in the empirical literature (that accounts for the endogeneity of dam location). While our results are in contrast with Duflo and Pande (2007) who find no effect of dams on malaria incidence, they fit the increasing body of anecdotal evidence suggesting dams have been associated with malaria (e.g. Kibret et al., 2019). Last, the coefficients for the lagged dependent variable are statistically significant and positive in all columns, and range from 0.5 to 0.6, suggesting some persistence of malaria.

Next, we examine how our findings change if we change the number of lags in the G2SIV estimation. Table 3 confirms our baseline findings the impact of dam construction on malaria in

the district in which a dam is located. According to columns (1)-(6), a one-dam increase in a district would increase malaria incidence by approximately 0.9 to 1.5 per cent. The impact on dam construction on malaria in downstream areas seems to be more sensitive to the number of lags used in the IV estimation. Columns 1 to 4 suggest there is no impact of dam on malaria in downstream areas. If we use further lags in columns 5 and 6, the estimates confirm our baseline results. As explained earlier, lag selection is chosen to minimize the variance of the standard error, so our baseline results are our preferred specification.

Alternative specifications

We test the robustness of our baseline findings to several alternative specifications. Table A2 in Appendix provides estimates from extended and reduced-form models. In the extended model, we include rainfall and temperature shocks, as is sometimes done in the literature. The use of reduced-form models allows examining whether bad controls drive our estimates. Across columns, the "dams" estimates are positive and statistically significant (except for the dam estimate in column 4 that is somewhat imprecise). The range of estimates implies a one-dam increase would increase malaria prevalence by about 0.5-1.5 per cent. Furthermore, when we look at the downstream effect of dams on malaria, we find evidence in support of our baseline results. In columns 1 and 2 (extended models), we find that dams do have a downstream effect on malaria. Similarly, the "upstream dams" estimates are positive, statistically significant, and are respectively 1.145 and 1.192 in columns 5 and 6.

Table A3 drops years in the estimation sample, as common factor approaches can be somewhat sensitive to sample selection. Columns 1 and 2 (3 and 4) exclude the first 3 (last 3) years of the dataset. In most cases, the estimates are in line with that of Table 2. They imply a one-dam

increase in the number of dams in a district would increase malaria prevalence by about 0.9-1.9 per cent. We find less consistent evidence for the downstream effect of dam on malaria. For instance, the "upstream dam" estimate in column 2 is however negative and statistically insignificant.

Table A4 uses 1 to 5-year lagged dams and upstream dams variables in the baseline model, instead of contemporaneous dam variables. This specification examines whether there are large lags in the effect of dam construction on malaria, especially in downstream areas. These estimates are particularly interesting because they confirm the malaria-increasing effect of dams persists over time in the district where the dam is located, as well as in downstream areas. In Table A5, we present estimation results relying on a slightly modified model than the one presented in Table 1. In particular, the number of common factors is equal to 4, following the selection criteria from Onatski (2010). Table A5 shows the main pattern of results remains the same. The dams coefficients remain positive, statistically significant, and in line with estimates from Table 2.

Agricultural mechanisms

What is the contribution of agricultural development to the increase in malaria incidence due to dam construction? Table 4 examines several pathways between dam and agricultural development. In particular, Table 4 estimates the impact of dam construction on several agricultural outcomes, i.e. agricultural production (i.e. value of the production of main crops, i.e. rice, wheat, sugarcane, jowar, bajra, maiz), yield of the main crops, gross irrigated area, gross cultivated area, and high yielding variety (HYV) crop cultivated area. Sample size vary across these outcomes but typically start earlier than 1975. More details can be found in the notes below Table 4.

Table 4. Agricultural mechanisms

Dependent variable	Agricultural production, log		Yield, log	Irrigata	Irrigated area, log		HVV are	HYV area, log	
variable			_						
	(1)	(2)	(3)	(4)	(5)	(6)	(7)	(8)	
Lagged									
dependent									
variable	0.435***	0.428***	0.105***	0.947***	0.547***	0.522***	0.381***	0.396***	
	(0.011)	(0.010)	(0.0066)	(0.037)	(0.001)	(0.010)	(0.010)	(0.008)	
Dams	0.232**	0.247**	0.198#	0.207**	0.314***	0.107***	0.022	0.036	
	(0.090)	(0.096)	(0.136)	(0.0834)	(0.051)	(0.030)	(0.090)	(0.082)	
Upstream dams	0.286***	0.361***	0.258**	-0.003	-0.045	-0.098***	0.566***	0.597***	
	(0.072)	(0.068)	(0.102)	(0.076)	(0.054)	(0.030)	(0.106)	(0.101)	
N N	5561	5,561	5,561	5280	3912	3902	3882	3882	
Number of districts Geographical	266	266	266	266	266	265	264	264	
controls	YES	YES	YES	YES	YES	YES	YES	YES	
Common factors	YES	YES	YES	YES	YES	YES	YES	YES	
District FE	YES	YES	YES	YES	YES	YES	YES	YES	
Rho	0.224	0.355	0.377	0.211	0.639	0.642	0.292	0.516	
p-val J test Number of	0.176	0642	0.075	0.440	0.403	0.298	0.725	0.578	
instruments Number of	120	120	120	40	140	140	140	140	
common factors	1	3	1	1	2	1	1	3	
Number of lags	5	5	5	1	6	6	6	6	

Notes: Robust standard errors in parentheses. *** p-val. <0.01, ** p-val. <0.05, * p-val. <0.1, # p-val. < 0.15. Dams and upstream dams coefficients are multiplied by 100. Rho is the fraction of the variance of the error term that is explained by the factor component. The sample includes 268 districts, instead of 358, due to missing data. Note these estimations are based on a sample ranging from 1966 to 1995, but the sample size varies across columns due to missing data. All G2SIV estimates. The sample includes 268 districts, instead of 358, due to missing data. All geographical variables (interacted with predicted dam incidence at the state level) from Table 1 are included except for the river gradient variables. Lag selection is determined to minimize the variance of the model's standard error.

First, we find that dam construction benefits agriculture in the areas in which they are built. For example, the "dams" estimate in columns 1 and 2 are positive and statistically significant. This suggests that one-dam increase in a district would increase agricultural production by about 0.2 per cent. The increase in agricultural production seems to be driven by the dam construction effect on irrigated and cultivated areas. A one-dam increase in a district would increase irrigated (cultivated) area by about 0.2-0.3 (0.1) per cent. Put simply, dams expand the role of irrigation as an agriculture practice in nearby farmlands in our context. The extent to which irrigation systems lack efficient drainage amplifies the existence of additional breeding ground for Anopheles mosquitoes.

These additional results portray a very different picture than the previous literature that often finds only negative or no effect on agriculture in the vicinity of dams (e.g. Strobl and Strobl, 2011; Blanc and Strobl, 2014; for a review, see Dillon and Fishman, 2019). In particular, our results directly contradict Duflo and Pande (2007) in the India context, but are also in line with more descriptive analyses (e.g. Singh, 1990; Pradhan and Srinivasan, 2022). To contextualize our results in the face of the literature, let's note that all the above-mentioned studies typically rely on state-year fixed effects in their modeling. We argued earlier why this modeling choice may lead to biased estimations.

In addition, in downstream areas, dam construction increases agricultural production (0.2-0.4%), yield (0.3%), and the cultivation of HYV crops (0.6%). The existence of positive benefits in downstream areas is in line with literature. However, we do not find that irrigated area or cultivated area¹⁰ respond to dam construction, so the benefits seem to be driven by changes in cropping patterns. In particular, the cultivation of HYV crops increases. These cropping patterns

¹⁰ The negative effect on cultivated area is economically small.

can arguably be conducive to the reproduction of mosquitoes and may explain the increase in malaria incidence in downstream areas.

Conclusions

The empirical literature on the links between large irrigation dams and malaria incidence mainly consists of case studies and anecdotal evidence. The few econometric studies regarding this topic remain vulnerable to several biases, among which the endogeneity bias due to the non-random nature of dam allocation and the omitted variable bias due to (unaccounted) time-varying heterogeneity.

Once we account for these biases, we find that large dams increase malaria incidence in districts where they are located, as well as in downstream areas. Our estimates imply a one-dam increase in the number of dams in the district increases annual malaria incidence by about 0.9 to 1.4 per cent, while the effect on malaria is about 1 to 1.5 per cent in downstream districts. This paper provides evidence that large irrigation dams increase malaria incidence and confirms an increasing literature based on case studies and anecdotal reports. We also find this malaria-increasing effect of dams lasts over time.

Furthermore, we examine the effect of dam construction on agricultural outcomes to highlight potential mechanisms between dam and malaria incidence. We find that dam construction benefits agriculture in the vicinity of dams, as well as in downstream areas. These positive effects are driven by increased irrigation in the vicinity of dams, while they are driven by changes in cropping patterns in downstream areas, where the cultivation of HYV crops (with larger water requirements) increases.

Both results related to malaria incidence and agricultural outcomes are novel in the empirical literature and should substantially alter future distributional cost-benefit analyses on large dam construction. They have clear implications for policymakers and researchers, especially within the context of the SDG agenda. From a policy standpoint, the construction of dams in malaria-sensitive regions should be coupled with targeted interventions, such as the wide deployment of insecticide-treated nets or the roll-out of future vaccines. From a methodological perspective, our paper highlights the role of time-varying heterogeneity in empirical studies analyzing the impact of irrigation infrastructures. Our results suggest the use of common factors approaches can be beneficial to researchers.

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Appendix

Table A1. Descriptive statistics of baseline sample

	Obs.	Mean	Std. dev.	Min.	Max.
Variables	(1)	(2)	(3)	(4)	(5)
Malaria incidence	2,845	3.725	6.711	0	95.51
Malaria incidence, log	2,845	0.0753	1.814	-4.605	4.559
Rainfall shock	2,012	0.0461	0.241	-0.773	1.281
Temperature shock	1,911	0.00473	0.311	-1.127	1.259
Number of dams in district	2,845	7.682	14.24	0	90
Number of dams upstream	2,845	13.53	27.99	0	246.0
Elevation 250-500m	2,845	0.846	1.636	0	7.348
Elevation 500-1000m	2,845	0.513	1.228	0	7.680
Elevation above 1000m	2,845	0.0483	0.263	0	2.806
Fraction of river gradient 1.5%-3%	2,845	0.193	0.301	0	1.893
Fraction of river gradient 3-6%	2,845	0.118	0.240	0	2.489
Fraction of river gradient above 6%	2,845	0.127	0.356	0	2.720
District area	2,845	22,100	30,552	0	288,978
River length	2,845	1,503	3,720	0	46,303
Fraction of district gradient 1.5%-3%	2,845	0.273	0.426	0	2.414
Fraction of district gradient 3%-6%	2,845	0.160	0.289	0	2.159
Fraction of district gradient above 6%	2,845	0.147	0.353	0	2.746

Table A2. Robustness analyses: alternative models

	Extende	ed model	Reduced form models			
Malaria, log	(1)	(2)	(3)	(4)	(5)	(6)
Malaria, log, t-1	0.541***	0.605***	0.542***	0.677***	0.508***	0.635***
	(0.003)	(0.002)	(0.010)	(0.718)	(0.009)	(0.006)
Rainfall shock	0.068***	0.039*				
	(0.022)	(0.020)				
Temperature shock	0.113***	0.069***				
	(0.010)	(0.017)				
Dams, own district	0.586**	0.865***	1.543***	0.486#		
	(0.256)	(0.255)	(0.434)	(0.326)		
Upstream dams	0.650***	0.821***			1.145**	1.192***
	(0.246)	(0.185)			(0.447)	(0.312)
N	1886	1886	2845	2845	2845	2845
Number of districts	241	241	355	355	355	355
Geographical controls	YES	YES	YES	YES	YES	YES
Common factors	YES	YES	YES	YES	YES	YES
District FE	YES	YES	YES	YES	YES	YES
Rho	0.192	0.430	0.202	0.388	0.221	0.404
J test	0.520	0.322	0.397	0.673	0.570	0.562
Number of instruments	154	154	126	126	126	126
Number of factors	1	3	1	3	1	3
Number of lags	6	6	6	6	6	6

Notes: Robust standard errors in parentheses. *** p-val. <0.01, ** p-val. <0.05, * p-val. <0.1, # p-val. <0.15. Dams and upstream dams coefficients are multiplied by 100. Rho is the fraction of the variance of the error term that is explained by the factor component. Note these estimations are based on a sample ranging from 1975 to 1995. All G2SIV estimates. The sample includes 355 districts, instead of 358, due to missing data. All geographical variables (interacted with predicted dam incidence at the state level) from Table 1 are included except for the river gradient variables.

Table A3. Robustness analyses: dropping years in the estimation sample

	1978-95	1978-95	1975-1992	1975-1992
Malaria, log	(1)	(2)	(3)	(4)
Malaria, log, t-1	0.422***	0.441***	0.524***	0.654***
	(0.002)	(0.000)	(0.009)	(0.006)
Dams	1.884***	0.992***	1.461***	0.913***
	(0.534)	(0.227)	(0.401)	(0.301)
Upstream dams	0.415*	-0.056	1.041**	1.085***
	(0.2176)	(0.090)	(0.430)	(0.309)
N	1907	1907	2845	2845
Number of districts	354	354	355	355
Geographical controls	YES	YES	YES	YES
Common factors	YES	YES	YES	YES
District FE	YES	YES	YES	YES
Rho	0.290	0.639	0.213	0.369
J test, p-val.	0.259	0.622	0.356	0.582
Number of instruments	140	140	140	140
Number of common factors	1	3	1	3
Number of lags	6	6	6	6

Notes: Robust standard errors in parentheses. *** p-val. <0.01, ** p-val. <0.05, * p-val. <0.1, # p-val. < 0.15. Dams and upstream dams coefficients are multiplied by 100. Rho is the fraction of the variance of the error term that is explained by the factor component. Note these estimations are based on a sample ranging from 1975 to 1995. All G2SIV estimates. The sample includes 355 districts, instead of 358, due to missing data. All geographical variables (interacted with predicted dam incidence at the state level) from Table 1 are included except for the river gradient variables. Columns 1 and 2 drop the first 3 years of the sample, while columns 3 and 4 drop the last 3 years of the sample.

Table A4. Robustness analyses: Lagging the dams coefficients

	t-	-1	t	-2	t-	-3	t-	-4	t-	-5
Malaria, log	(1)	(2)	(3)	(4)	(5)	(6)	(7)	(8)	(9)	(10)
Malaria, log	0.699***	0.562***	0.717***	0.571***	0.531***	0.484***	0.553***	0.579***	0.676***	0.670***
	(0.009)	(0.014)	(0.015)	(0.018)	(0.019)	(0.032)	(0.020)	(0.044)	(0.074)	(0.089)
Dams, lagged	0.589#	1.779***	0.746#	1.193#	2.264***	0.030***	1.693***	1.342*	0.121	0.894*
	(0.401)	(0.458)	(0.499)	(0.7581)	(0.419)	(0.009)	(0.398)	(0.725)	(0.419)	(0.459)
Upstream dams, lagged	0.715**	0.836*	0.978**	0.823*	1.294***	0.021***	0.795#	1.612**	0.980**	1.413**
	(0.285)	(0.446)	(0.380)	(0.499)	(0.446)	(0.006)	(0.494)	(0.663)	(0.406)	(0.628)
N	2,845	2,845	2,845	2,845	2,483	3,144	2,141	2,798	3,115	3,778
Number of districts	355	355	355	355	353	356	352	354	355	356
Geographical controls	YES									
Common factors	YES									
District FE	YES									
Rho	0.387	0.201	0.383	0.200	0.447	0.231	0.514	0.169	0.373	0.148
p-val J test	0.632	0.560	0.559	0.266	0.418	0.254	0.663	0.319	0.479	0.405
Number of instruments	120	120	100	100	100	80	100	80	60	40
Number of common factors	3	1	3	1	3	1	3	1	3	1
Number of lags	5	5	4	4	4	3	4	3	2	1

Notes: Robust standard errors in parentheses. *** p-val. <0.01, ** p-val. <0.05, * p-val. <0.1, # p-val. <0.15. Dams and upstream dams coefficients are multiplied by 100. Rho is the fraction of the variance of the error term that is explained by the factor component. Note these estimations are based on a sample ranging from 1975 to 1995. All G2SIV estimates. The sample includes 355 districts, instead of 358, due to missing data. All geographical variables (interacted with predicted dam incidence at the state level) from Table 1 are included except for the river gradient variables. Dams coefficients are replaced by their lagged values. For example, columns 1 and 2 lag the dams coefficients by one period, while column 9 and 10 lag the dams coefficients by 5 periods.

Table A5. Robustness analyses: baseline results based on different estimation than in Table 1 (change in common factors to 4)

Malaria, log	(1)	(2)
Molaria log t 1	0.646***	0.514***
Malaria, log, t-1		
	(0.006)	(0.009)
Dams	0.853***	1.350***
	(0.307)	(0.401)
Upstream dams	1.106***	0.946**
	(0.309)	(0.420)
N	2,845	2,845
Number of districts	355	355
Geographical controls	YES	YES
Common factors	YES	YES
District FE	YES	YES
Rho	0.398	0.385
J test	0.555	0.218
Number of instruments	140	140
Number of common factors	3	1
Number of lags	6	6

Notes: Robust standard errors in parentheses. *** p-val. <0.01, ** p-val. <0.05, * p-val. <0.1, # p-val. < 0.15. Dams and upstream dams coefficients are multiplied by 100. Rho is the fraction of the variance of the error term that is explained by the factor component. Note these estimations are based on a sample ranging from 1975 to 1995. All G2SIV estimates. The sample includes 355 districts, instead of 358, due to missing data. All geographical variables (interacted with predicted dam incidence at the state level) from Table 1 are included except for the river gradient variables.

Description of agricultural outcomes used in Table 4 (source: Duflo and Pande, 2007)

Agricultural production: the value of the production of main crops, includes rice wheat, sugarcane, jowar (sorghum), bajra (pearl millet), maize in rupees per thousand tons (average prices for crops 1960-65 are used).

Yield: yield of the main crops: rice wheat, sugarcane, jowar (sorghum), bajra (pearl millet), maize, the yield being measured as the number of rupees per hectare (average prices for crops 1960-65 are used).

Gross irrigated area: gross irrigation area, in thousand hectares.

Gross cultivated area: gross cultivation area being measured in thousands hectares.

HYV area: area under HYV crop cultivation