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Applying the Miceli Model to Explain Cooperation in Municipal Solid Waste Management

Kelly J. Tiller and Paul M. Jakus

As traditional methods of municipal solid waste management (MSWM) become increasingly expensive due to increased regulation, many local governments are considering cooperation as a waste management strategy. A theoretical model is used to specify a partial observability probability model to analyze the decision Tennessee counties made to form either a singlecounty solid waste region or a multi-county region. We find that, while economies of scale may be a factor in the consolidation decision, similarities and differences between counties in current individual provision levels of solid waste services, ability to pay for services, and expectations for future solid waste service demands are statistically more important.

Key Words: regional cooperation, municipal solid waste, waste management, regionalization

Municipal solid waste (MSW) refers to a specific portion of the waste stream generated by residential, commercial, institutional, and some industrial sources, primarily solid waste.¹ Traditionally, municipal solid waste management (MSWM) has been the responsibility of local governments, with landfilling the most common method of disposal. In the early 1990s, federal regulations affecting traditional methods of solid waste disposal increased the cost by as much as five- to tenfold. In addition, the vast majority of states passed recycling laws, or adopted recycling, diversion, or waste reduction goals, and many states approved comprehensive waste management legislation requiring long-term planning (Steuteville 1995). As MSWM has grown increasingly complex and expensive, one strategy that some communities have developed to meet new MSWM challenges is regional (e.g., multi-county, multi-community) cooperation. Cooperation is a process whereby

neighboring cities, counties, or other governmental entities pool resources to address local challenges, taking advantage of the potential economies of scale associated with many aspects of MSWM. Many states have also included incentives, provisions, and/or mandates for formation of solid waste regions as an element of MSWM legislation.

Beyond the narrow arena of solid waste management, rural regions are faced with ever-tightening budgetary environments and must investigate alternative means to supply necessary or mandated public goods and services. A common method is to exploit economies of scale by merging or consolidating service regions for public goods. Following Gyimah-Brempong's (1987) pioneering empirical work on consolidation of law enforcement agencies, many researchers have applied a translog cost function approach to evaluating scale efficiencies in the consolidation of rural school districts [see, among many studies, the recent analysis of rural Arkansas school districts by Garrett and Dodson (2004)]. Other applications to provision of public goods in rural regions include studies of county-level extension services by Garrett (2001) and rural roads by Deller and Nelson (1991).² Despite its obvious

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¹ Not included are materials that also may be disposed of in landfills, but are not generally considered MSW, such as construction and demolition debris, municipal wastewater treatment sludge, and nonhazardous industrial waste (U.S. Environmental Protection Agency 2003).

² Rather than the "standard" cost function approach used by many, Deller and Nelson (1991) used a Farrell frontier model to evaluate efficiencies associated with consolidation.

appeal, the cost function approach often presents empirical difficulties in that one must have explicit measures for both inputs and outputs for public goods. Inputs and outputs may not be easily quantified, and a review of the literature reveals some degree of anguish on the part of researchers with respect to this issue (e.g., Garrett 2001, p. 816). Even if one has reasonably good measures for inputs and outputs, one must often assume away difficulties associated with jointness in production of outputs.

Further, it is not just scale economies that matter in the consolidation decision. A local government may enjoy scale economies of a merger in the provision of a public good yet choose not to take advantage of economies because the joint level of provision is not an optimum for the entity. For example, Jacques, Brorsen, and Richter (2000) show that rural Oklahoma schools can achieve scale economies with larger school districts but that student achievement declines as districts get larger. Given this tradeoff, a community may rationally reject a cost-saving merger if the jointly provided public good (student quality) is unsatisfactory.

Our empirical analysis concerns county-level cooperation decisions made following passage of the 1991 Tennessee Solid Waste Management Act, which provided a rich data resource for quantitatively analyzing cooperation decisions. The theoretical approach follows Miceli's (1993) model as developed to address public school district consolidation. Miceli's model allowed for mergers only if both scale economies and the joint level of public goods provision represented a Pareto improvement for all. Similar to DeBoer (1995), we use the Miceli model to evaluate consolidation of solid waste management districts, but our analysis differs from DeBoer's in that we use Poirier's (1980) partial observability approach for model estimation. While DeBoer's approach looked at characteristics of only one potential partner in a consolidation decision, this model recognizes that the decision to merge districts is a jointly determined outcome that is the result of independent decisions of individual districts. This mating game approach allows consideration of characteristics and differences of all potential parties in the agreement, an important factor not addressed in DeBoer's study. The partial observability approach offers advantages over the cost function approach because it is wellgrounded in economic theory yet far less dataintensive.

We first review Miceli's theoretical model explaining the joint provision of a public good, where the model is used to specify the factors important for an empirical test. The partial observability model is then reviewed, followed by discussion of the test data and presentation of the empirical results. We conclude with an agenda of future research.

Economies of Scale in a Model of Regional Cooperation

Cooperation, or consolidation, in the provision of public goods was explicitly expressed in Miceli's (1993) version of the Tiebout model. The model recognizes that public goods, such as the provision of solid waste services, are funded out of property tax revenues. Following Miceli's notation, a budget constraint for a member of county *i* can be written as

$$y_i = x_i + p(1+t_i)h_i,$$

where y_i is income, x_i is a numeraire, p is the price of housing, t_i is the property tax rate, and h_i is the quantity of housing. Given this income constraint, one can optimize utility and specify an indirect utility function,

$$V = V(y_i, p(1+t_i), g_i),$$

where g_i is the level of public goods provided by the local government. According to the Tiebout hypothesis, members of a county will choose a residence so as to maximize this utility function on the basis of the cost and provision of public goods. The county's tax base is given by

$$B_i = pH_i + S_i,$$

with H_i being the total housing stock in county *i* and S_i being the value of non-residential taxable property in the county.

Assume that MSWM is the sole public good provided by the county. Let $c(n_i)$ denote the unit cost function for providing MSWM to *n* residents of county *i*. As the derivative of the unit cost function is negative or positive, marginal costs

are decreasing or increasing.³ Economies of scale exist if the marginal cost, dc/dn_i , is less than the average unit cost, $c(n_i)/n_i$. Diseconomies occur if $dc/dn_i > c(n_i)/n_i$. A balanced budget for the county is then given by⁴

$$t_i B_i = c(n_i) g_i \, .$$

Dividing both sides by n_i and rearranging yields

(1)
$$t_i = \frac{\frac{c(n_i)}{n_i} \times g_i}{\frac{B_i / n_i}{B_i / n_i}}$$

Equation (1) shows the supply of the public good in t-g space, where the slope of the supply function is defined by the per capita unit cost of provision (in the numerator) and the per capita tax base (denominator).

Assume now that a multi-county MSWM region is proposed. Such a regional administration provides solid waste services g_R , which may differ from g_i . Assume further that administrative costs are shared in proportion to county population. Total costs to county *i* are now given by $(n_i/n)[c(n)g_R]$, where *n* is the regional population. The balanced budget supply of solid waste services under regionalization, t_i^R , is given by

(2)
$$t_i^R = \frac{\frac{c(n)}{n} \times g_R}{\frac{B_i}{n_i}}.$$

The average unit cost of solid waste services, c(n)/n, will fall if there are unexploited economies of scale associated with regionalization. If scale economies exist, the same level of solid waste services may be provided at a lower cost, yielding a lower tax rate for the community.

Miceli (1993, p. 351) also notes that two counties may be currently providing different levels of service, say g_i and g_j , and that the jointly feasible level of provision, g_R , may differ from the initial amount offered by either community. Even if scale economies exist, each community must decide if the change in the level of provision is worth the change in the community tax rate. This suggests that, in addition to economies of scale measures, current levels and future levels of solid waste services will be considered by entities. Finally, Brasington's (1999) study of school district consolidation in Ohio notes that, given the relationships in equations (1) and (2), "communities rich in property value will not be inclined to merge with property-poor communities unless they are sufficiently compensated by cost savings" (p. 378).⁵

Municipal Solid Waste Management in Tennessee

The state of waste management in Tennessee appears to be representative of MSWM in the United States. According to Kaufman et al. (2004), annual MSW generation in Tennessee is 1.27 tons per person, compared to a national average of 1.31 tons per person (the range nationwide is from 0.68 to 1.73 tons per person). Recycling rates in Tennessee are estimated to be 26.4 percent compared to a national average of 26.7 percent (the range nationwide is 1.0 percent to 48.8 percent). Tennessee landfills 71.6 percent of all waste generated, compared to a national average of 65.6 percent.

Similar to many states, Tennessee passed a comprehensive Solid Waste Management Act in 1991. Passage of this Act was the first effort by the state to require all counties to meet a minimum standard level of service in the area of solid waste management. Elements of the legislation addressed solid waste planning, collection, disposal, recycling, education, and funding, as well as collection and disposal of problem wastes.

Specifically, the legislation required each county to form a solid waste region and to develop a ten-year solid waste plan for the region. The legislation included a number of required elements, including the requirement that at least 90 percent of all residents in the region have collection service available to them.⁶ The minimum collection service level was established to be a

³ If dc/dn = 0, then there are no scale economies, and $c(n_i) = c(n_i+1)$.

⁴ If exogenous planning and operating funds for an MSWM district are given by G_{i} , total revenue available for solid waste management is given by $t_iB_i + G_i$. Because planning funds were provided on only a one-time basis, they are ignored in this analysis.

⁵ See also Brasington (2003).

⁶ Another requirement was that all counties were to reduce the amount of MSW entering landfills or incinerators by 25 percent over a four-year period.

system of drop-off convenience centers for garbage collection. Additionally, each county had to establish a minimum of one collection center for recyclable materials. Grants were provided for planning purposes, but not for ongoing operational costs.

Counties were permitted to form multi-county solid waste regions or a single-county region. According to Section 12.a.2 of the Act, "The preferred organization of the regions shall be multi-county. Any county adopting a resolution establishing a single-county region shall state the reasons for acting alone in the resolution." No upper limits were placed on region size, provided that all region members were contiguous counties.⁷

Analysis conducted in 1991 by the University of Tennessee Waste Management Research and Education Institute (Barkenbus et al. 1991) indicates that potential scale economies exist in Tennessee (Figure 1). Savings are due primarily to declining average costs of landfilling in a Subtitle D-compliant landfill up to an efficient tonnage level. Economies of scale at landfills are based on (i) tonnage per day received at the facilities, (ii) compaction rates achieved, as measured by inplace refuse densities, (iii) percentage of landfill

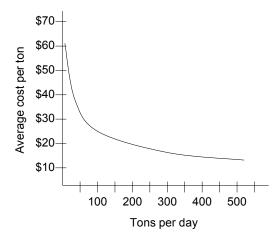


Figure 1. Average Cost per Ton of Landfill Waste in Tennessee

Source: Barkenbus et al. (1991)

volume taken up by dirt required for various cover operations, and (iv) average height of refuse over the liner (County Technical Assistance Service 1991). The cost savings available to larger facilities are due to the fact that more waste can be handled with relatively small increases in equipment and labor, and that there is an inverse relationship between the tonnage received per day and percentage dirt required for cover. Further scale economies may be captured at the collection stage as well as at the disposal stage.

In response to the 1991 Tennessee Solid Waste Management Act, 45 of Tennessee's 95 counties joined multi-county solid waste regions in 1993. In addition to the 50 one-county regions, one 2county region, seven 3-county regions, three 4county regions, and one 10-county region were formed (Figure 2). The decision each county made regarding the formation of a solid waste region provides a natural experiment for testing the Miceli model.

Methods and Data

Econometric Methods

The theoretical model suggests that in addition to the potential to exploit economies of scale, major factors affecting the cooperation decision are per capita property values, population, current levels of service, expected future levels of service, and differences in these measures. Unfortunately, we cannot directly measure the "desire" of a county to join a region. This is because the observed outcome—joining a multi-county region or not—is the result of an agreement between two entities, not one. Thus, the appropriate method of modeling the outcome is a "partial observability" model (Poirier 1980).

Consider the desire by county 1 to join county 2 as measured by the latent variable y_1^* , and parameterized according to $y_1^* = f(x_1; \beta)$, where the vector of explanatory variables x_1 is given by the theoretical model. The desire of county 2 to join county 1 is measured by the latent variable y_2^* and parameterized by $y_2^* = g(x_2; \beta)$. Following the standard random utility model hypothesis, we observe county 1 desiring to join county 2, denoted as $y_1 = 1$, if $y_1^* > 0$.⁸ If $y_1^* < 0$, then the

⁷ Tennessee's 95 counties are divided into nine multi-county Development Districts, voluntary associations of municipal and county governments that provide members with assistance in addressing economic development and growth. There do not, however, appear to be other public goods or services in Tennessee for which state legislation requires or even strongly encourages multi-county cooperation.

⁸ The random utility model asserts that county 1 will desire to cooperate with county 2 if the utility of the regional partnership exceeds the utility of the single county MSWM "region", or $U^{R}(t_{1}^{R}, g_{R}) > U^{1}(t_{1}, g_{1})$.

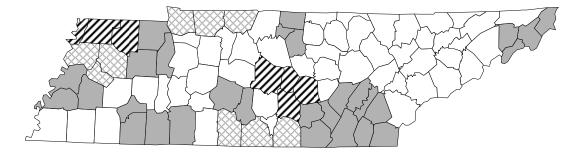


Figure 2. Solid Waste Management Regions in Tennessee

county does not wish to join, and $y_1 = 0$. Similarly, county 2 will desire to form a regional partnership with county 1, denoted as $y_2 = 1$, if $y_2^* > 0$, with $y_2 = 0$ otherwise. The analyst does not observe the latent variables y_1^* or y_2^* . Given the fact that both entities must agree to form a partnership, y_1 and y_2 are not observed either. Instead, what is observed is the joint outcome, $z = y_1 \times y_2$. The observed joint outcome, z, will take the value of one (an agreement) if and only if $y_1 = y_2 = 1$. If either county chooses not to cooperate, z = 0. Poirier (1980) terms this a partial observability model, which can be modeled as a bivariate probit with the likelihood function

$$\ln L = \sum_{z=1} (\ln \Phi_2 [\beta x_1, \beta x_2, \rho]) + \sum_{z=0} \ln(1 - \Phi_2 [\beta x_1, \beta x_2, \rho]),$$

where $\Phi_2(\bullet)$ is the bivariate normal distribution and ρ is the correlation between the two entities' choices.

Data

County-level data were available from a variety of sources. MSWM regional status data were provided by the Tennessee Department of Environment and Conservation, Division of Solid Waste Assistance. The same department also provided information on the current state of solid waste collection in each county, including the presence of landfills and the percentage of unmanaged waste in a county. Population density and population growth rates were gathered from the U.S. Bureau of the Census, while property tax base data were found in the Tennessee Statistical Abstract. Means for the single-county and multicounty regions are shown in Table 1.

Following Brasington's (1999) data arrangement method, the 95 counties in Tennessee had 367 potential cooperative regional partnerships in the provision of solid waste services.⁹ For any given county, a potential partner may consist of one or more counties as long as the first county is contiguous with the potential partner. The data are arranged randomly in that assigning a particular county or group of counties to position one or position two in the data set did not intentionally follow any pattern. Further, given that potential partners may consist of more than one county, we now refer to potential partners as "entities" or "units".

An entity's per capita assessed valuation measures the denominator in equation (2), while the population of the entity proxies for average unit cost of solid waste services (the numerator). Following Brasington (1999), we anticipate that differences in assessed property valuation cause entities to be less likely to form a solid waste region. Economies of scale are measured using two different variables. The first method follows Callan and Thomas (2001) and uses population density as a proxy for scale economies, whereas the second method follows Brasington (1999) and uses population as a proxy. The economies of scale hypothesis is supported with a negative coefficient on the linear population density (population) term and a positive sign on the square root of population density (population) term.¹⁰ It is not clear that differences in population density (population) across entities would be positively or negatively related to scale economies.

⁹ See Brasington's (1999) Appendix A, p. 391. Prior to conducting any econometric analysis, Dr. Brasington was gracious enough to review our data arrangement.

¹⁰ We departed from the standard "squared" non-linear term to avoid scaling issues in the maximum likelihood estimation. The negative sign on the linear term and a positive sign on the square root non-linear term will yield the familiar inverted-U shape for a function.

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Variable	Mean	Standard deviation	Minimum	Maximum
Population density (persons/square mile)				
45 "join" counties	79.25	82.58	18.94	495.93
50 "not join" counties	126.69	207.50	15.63	1,054.41
Per capita assessed value (\$1000)				
45 "join" counties	6.68	1.49	4.89	11.06
50 "not join" counties	7.35	2.66	4.43	16.76
Subtitle D landfill ($0 = no \ access, \ 1 = access$)				
45 "join" counties	0.31	0.47	0	1
50 "not join" counties	0.54	0.50	0	1
Percentage waste unmanaged				
45 "join" counties	35.9	22.5	0	77.1
50 "not join" counties	26.4	24.8	0	78.0
Percentage population growth rate				
45 "join" counties	2.73	8.27	-10.50	41.10
50 "not join" counties	5.94	8.88	-5.70	39.40

Table 1. Variable Means, Standard Deviations, and Minimum and Maximum Values

Current levels of solid waste services (g_i) are measured in two ways. First, the presence of a Subtitle D-compliant landfill operated by the county or contractually available to the county is a measure of current services. Secondly, the Tennessee legislation mandates that 90 percent of a region's residents must have access to some form of solid waste collection, a minimum level of recycling opportunity, and a 10-year assurance of disposal capacity. In essence, the legislation mandates a minimum level of g_R . Some counties satisfied all of these requirements prior to the legislation passage $(g_i \ge g_R)$, whereas other counties did not satisfy any $(g_i < g_R)$. We capture this legislative influence with a variable measuring the percentage of unmanaged waste in a county at the time the legislation was passed. Counties with higher percentages of unmanaged waste have further to go to meet state-mandated service level requirements. It is hypothesized that the more effort required on the part of a county to meet state-mandated requirements, the more likely it will be to join a multi-county solid waste region to achieve g_R . That is, the net marginal benefits of cooperation are likely to be higher for counties with infant solid waste management programs than those with well-developed programs. The difference in the percentage of unmanaged waste represents a difference in the level of current service offered by each entity, $g_i - g_j$. It is hypothesized that having greater differences in current service levels reduces the likelihood of cooperation.

Econometric Results

Three econometric specifications were tested (Table 2). The first specification focuses only on those variables that capture the economies of scale hypothesis, the current provision of solid waste services, and a measure of the specific aspects of the legislation providing the impetus for consolidation. In Model 1 of Table 2, the linear population density term is statistically insignificant, with a P-value of 0.17, whereas the non-linear term is significant. These results do not clearly support the scale economies hypothesis regarding the decision to form a solid waste region. In contrast, the difference in per capita assessed valuation is statistically significant. This suggests that the greater the relative disparity in county wealth, the less likely the entities are to form a solid waste region. Access to a Subtitle D landfill also makes the entity less likely to form a multi-county region. We interpret this result as finding that entities that satisfy one of the minimum legislative

	Model 1	1	Model 2	12	Model 3	lel 3
Variable	Beta	t-stat	Beta	t-stat	Beta	t-stat
Intercept	-2.749	-0.530	-3.454	-0.622	-2.334	-0.418
Population Density	-0.373	-1.388	-0.551	-1.612		
Square Root of Population Density	1.171^{b}	1.733	1.676^{a}	2.015		
Population					-0.479	-1.495
Square Root of Population					0.455 ^b	1.727
Difference in Population Density	0.053	0.271	0.105	0.560		
Difference in Population					0.004	0.227
Per Capita Assessed Valuation	-0.208	-0.281	-0.132	-0.167	-0.047	-0.059
Square Root PC Assessed Valuation	1.441	0.364	1.455	0.345	-0.884	0.212
Difference in PC Assessed Valuation	-0.357 ^a	-2.691	-0.442 ^a	-3.355	-0.412 ^a	-2.940
Subtitle D Landfill	-0.355 ^a	-2.209	-0.400 ^a	-2.624	-0.425 ^a	-1.980
% of Waste Unmanaged	1.276 ^a	3.941	1.299 ^a	3.744	1.453 ^a	2.341
Difference in % of Waste Unmanaged	-3.696 ^a	-5.408	-3.893 ^a	-5.81	-3.838 ^a	-5.286
Population Growth Rate			-0.022 ^a	-2.842	-0.024 ^a	-2.129
Difference in Population Growth Rate			-0.003	-0.953	-0.005	-1.129
β	-0.989 ^a	-8.075	-0.997 ^a	-8.982	-0.875 ^b	-1.740
Log-likelihood	-178.851	51	-172.991	91	-177.000	000
Chi-square	57.770 ^a) ^a	69.496 ^a	16 ^a	61.4	61.488^{a}
% correct	32.4%	\ 0	44.4%	%	56.4	56.4%
% "not join" correct	9.6%		32.5%	%	60.2%	2%
% "join" correct	98.9%	0	79.6%	%	45.2%	2%

requirements (i.e., those for which $g_i \ge g_R$ prior to the legislation) are less likely to find formation of a multi-county solid waste region an improvement. Finally, as the percentage of unmanaged waste in a county increases, the greater the likelihood of a regional partnership. This tendency is tempered by the negative effect of the difference between the percentages of unmanaged waste: partnerships are made between those with similar unmanaged waste problems. The correlation coefficient, p, is statistically significant, indicating that the decisions of the two entities are "connected" and that the bivariate approach correctly accounts for this dependence across entities. This specification did an excellent job of predicting those counties that would join a multi-county solid waste region, but predicted rather poorly those that would not join a region (fewer than 10 percent correctly predicted).

To improve the predictive capability of the model, we consider another potentially important aspect of the legislation: the 10-year assurance of disposal. This suggests that a measure of future growth in solid waste generation should be reflected in the model. The second specification reflects this aspect by adding the population growth rate to the model (Model 2). In this case, the population density terms are both statistically significant if one chooses a P-value of 0.11. Larger differences across entities in per capita assessed valuation make cooperation in solid waste management less likely. The presence of a Subtitle D landfill also reduces the probability of a cooperative arrangement. Increasing amounts of unmanaged waste lead to cooperation but, again, only among those entities sharing similar levels of unmanaged waste. Finally, because those units with higher population growth rates will be generating an ever greater quantity of solid waste, high growth rates reduce the probability that an entity will join a multi-county solid waste region. This model maintains the excellent prediction record of Model 1 for those choosing to join a region (almost 80 percent correctly predicted), while greatly improving the predictive record for those not choosing to join (with 32 percent correctly predicted).

Finally, Model 3 replaces the population density terms with measures of population [the measure used by Brasington (1999) and others]. Similar to Models 1 and 2, this specification provides relatively weak support for the economies of scale hypothesis. The linear population term is statistically insignificant at conventional levels, whereas the non-linear term is significant at the 10 percent level. All other variables retain signs and levels of significance similar to those of the two initial specifications. This model did the best job at correctly predicting the outcome of the decision process, with an overall success rate of 56.4 percent.

Conclusions and Policy Implications

The partial observability approach to modeling consolidation decisions has been shown to be a useful and relatively simple analytical method that may prove to be of interest to other researchers. Similar to the cost function framework, the partial observability model can be well grounded in economic theory and yet avoid many of the data complications of the former. The Miceli theoretical model of consolidation provides a clear set of testable hypotheses and can be readily implemented in the empirical framework offered by the partial observability approach.

With respect to our empirical application, we find some support for economies of scale in the formation of solid waste regions in Tennessee, but the evidence is not overwhelming. While the coefficients indicate that scale economies are likely to be present, only one of these coefficients was consistently significant across specifications. Instead, the statistically strongest factors in the empirical model proved to be related to current and future levels of solid waste services. Access to a Subtitle D-compliant landfill and low levels of unmanaged waste for an entity made that entity less likely to join a multi-county solid waste region relative to those without access to a landfill and with high levels of unmanaged waste. Further, those entities with high future growth in solid waste generation were less likely to join in multi-county regions. Taken collectively, the statistical results highlight Miceli's point that the existence of scale economies is a necessary but not sufficient condition for a merger between two entities. Indeed, our results suggest that the joint provision level and differences in current individual provision levels are the driving forces in the decision of Tennessee counties of whether to join a multi-county solid waste region.

Tiller and Jakus

While the results presented in this paper are satisfying, the statistical models do not include other important factors that are difficult to measure. Tiller (1996) argues that political risk and loss of local autonomy, for example, are important contributors to the cooperation decision. Dinar and Wolf (1997) echo this argument, finding that political considerations are the stabilizing influence in regional solutions that are economically feasible. Tiller's (2001) case studies examining MSWM cooperation in Tennessee suggest that factors difficult to quantify, such as leadership qualities, personal relationships among leaders within and across counties, and historical experiences, also play an important role. Future research on cooperative outcomes should endeavor to incorporate these factors into the analysis. Another alternative to multijurisdictional cooperation that many local governments are also considering, particularly as budget pressures mount, is privatization in the provision of some goods and services. Future research that incorporates contrast and comparison of cooperation decisions and privatization decisions would further contribute to the understanding of local entity decision processes.

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